Regional characterisation of hard-bottom nursery habitat for juvenile Caribbean spiny lobster (Panulirus argus) using rapid assessment techniques

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Abstract    Shallow, hard-bottom habitat constitutes approximately 30% of the coastal waters of south Florida, United States, yet it is a chronically understudied feature of the marine seascape in this region. In this study, we characterised the general biogeographic and structural features of shallow benthic hard-bottom communities in the Florida Keys, and related those to the abundance of juvenile Caribbean spiny lobster (Panulirus argus), the target of one of Florida’s most economically valuable fisheries. We used rapid assessment techniques to survey more than 100 hard-bottom sites in the Florida Keys to estimate the percentage bottom coverage of vegetation (seagrass and macroalgae) and the abundance of sponges, octocorals, hard corals, and other crevice-bearing structures, as well as the abundance of juvenile lobsters. Using a multivariate statistical approach, we evaluated the relationship between habitat and size-specific juvenile lobster abundance and quantitatively verified the existence of six generally accepted biogeographic subregions. Although the types of hard-bottom shelters used by juvenile lobsters varied somewhat among these subregions, in all regions, branching-candle sponges and octocorals were under-used by lobsters, whereas loggerhead sponges, coral heads, and solution holes were over-used (i.e., used more frequently than expected based on their availability). There was also an ontogenetic transition in the shelter preference of juvenile lobsters; small juveniles tended to occupy a variety of sponges, whereas large juveniles preferred hard structures such as coral heads and solution holes. This study yields the first quantitative biogeographic description of hard-bottom communities of the Florida Keys, and confirms the suspected relationship between the structural features of hard-bottom habitat and the value of these communities as nurseries for juvenile spiny lobster.

Keywords    Florida Keys; habitat use; essential habitat

INTRODUCTION

Hard-bottom habitat is a ubiquitous feature of the Florida Keys, Florida, United States marine ecosystem, covering over 30% of the region’s shallow coastal seafloor (Zieman et al. 1989, Herrnkind et al. 1997). Hard-bottom occurs in shallow (<5 m) coastal waters and is characterised by low relief (<0.5 m), limestone bedrock overlain by a thin veneer of sediment and populated by a complex assemblage of sponges, octocorals, corals, and macroalgae (Chiappono 1996). It has been variously referred to as live-bottom habitat, octocoral-dominated hardgrounds, exposed bedrock outcrops, alcyonarian-sponge communities, nearshore rock-bottom habitat, and algal-dominated hard-bottom (see Schomer & Drew 1982; Chiappono 1996). In the shallow waters surrounding the Florida Keys
off the southern tip of Florida, hard-bottom habitat intersperses with sand, calcareous mud-bottom, and seagrass meadows (primarily *Thalassia testudinum*, *Syringodium filiforme*, and *Halodule wrightii*).

A large but unknown number of motile macrofaunal species dwell in hard-bottom areas of the Florida Keys. Many species use hard-bottom habitat opportunistically (e.g., bonefish, tarpon, sharks, turtles), whereas others are obligate dwellers of hard-bottom and are rarely found elsewhere (e.g., octocorals, many anemones, spider crabs, octopus, and certain fishes; Lindeman & Snyder 1999). A number of fishes use hard-bottom as a nursery before migrating to other habitats, such as coral reefs (Baron et al. 2004). The Caribbean spiny lobster (*Panulirus argus*, Latreille, 1804), which sustains one of the most economically important fisheries in Florida (Hunt 2001), also uses the hard-bottom habitat as a nursery (Herrnkind et al. 1997).

Following an extended period of larval dispersal in the open sea, Caribbean spiny lobster postlarvae arrive inshore every month on new moon flood tides (Acosta et al. 1997), seeking structurally complex vegetation in which to settle. In south Florida, they preferentially settle in clumps of red macroalgae within hard-bottom habitat (Marx & Herrnkind 1985; Herrnkind & Butler 1986; Herrnkind et al. 1994; Butler et al. 1997), but highly-fouled mangrove roots (Marx 1986; Acosta & Butler 1997), seagrass (Herrnkind et al. 1994; Acosta & Butler 1997) and possibly other complex substrates may also harbour newly settled lobsters in other regions (Briones-Fourzan & Lozano-Alvarez 2001). After a few months, juveniles emerge from their vegetated settlement sites and take up daytime refuge under hard-bottom structures (e.g., sponges, solution holes, corals) that provide them with suitable shelter from predators (Eggleston et al. 1990; Forcucci et al. 1994;
Butler et al. 1995). Later, as subadults, they migrate to coral reefs to join the adult population (Lyons & Kennedy 1981). To complete their complex life cycle, lobsters therefore use most of the region’s marine habitats beginning with the hard-bottom nursery.

Characterising the regional biogeography of hard-bottom habitat and understanding the interaction between shelter availability and its usage by lobsters may help us determine the regions and features of nursery habitat most critical to juvenile lobster. We therefore surveyed juvenile lobster abundance and selected structural features of hard-bottom habitat at more than 100 sites throughout Florida Keys to detail regional patterns in hard-bottom habitat structure, juvenile lobster abundance, and their interrelationship.

### MATERIALS AND METHODS

#### Field surveys

Field surveys of benthic habitat and surface visual assessments were conducted at 381 sites throughout the Florida Keys during May–July of 1993 and 1994 (Fig. 1). In 1993, we used SCUBA to survey the benthos and lobsters at 26 sites, whereas in 1994 we assessed the habitat types of 355 sites, 89 of which included SCUBA-based benthos and lobster abundance estimates, and the remainder were visual assessments from the surface. When this study began, GIS-based benthic habitat type maps were not yet available for use in our selection of survey sites, so survey sites could not be stratified by habitat. Also, the distribution of various shallow benthic habitats was unknown. Therefore, we used a uniform sampling protocol wherein the basic criterion guiding our selection of survey locations was to maintain a more-or-less even separation of one nautical mile (1.85 km) between sites along transect lines running approximately north-south. However, in some areas, the complex topography of islands and shallow water in the Florida Keys precluded exact north-south transects and one nautical mile inter-site distances (Fig. 1).

Depending on the initial characterisation of bottom type, up to four types of data were collected at each site. Upon arrival at each pre-determined site location, we first noted our exact position (latitude and longitude) and then characterised the site into one of three broad habitat types (sand, seagrass, or hard-bottom) based on visual inspection of the bottom. If the site was predominantly sand or seagrass, no further data were collected. If the site was hard-bottom and thus contained large sponges, octocorals, or other potential juvenile lobster shelters, we then proceeded with quantitative surveys conducted by divers. These included surveys of the abundance of: (1) general vegetation types (i.e., seagrass, red macroalgae, and green algae) that influence lobster settlement; (2) prominent benthic structural features that are potential lobster shelters (i.e., sponges, corals, solution holes); and (3) juvenile lobsters. At each dive site, percentage cover of bottom vegetation was determined from three, non-overlapping 25 m long tape transects (four transects in 1993) placed haphazardly along the bottom at each site. Transects containing more than 20% seagrass were rejected and repositioned. Divers travelled along each tape and recorded the size and positions of all patches of seagrass, green macroalgae (e.g., Halimeda spp., Penicillus spp.) and red macroalgae ( Laurencia spp.) larger than 5 cm.

Sponges, octocorals, and other crevice-bearing structures large enough to be potential shelter for juvenile lobster (Childress 1995) were surveyed using three 2 m × 25 m belt transects using the same transect positions as for the vegetation surveys. Within each 50 m² belt transect, divers enumerated all of the loggerhead sponges (Speciospongia vesparium), vase sponges (Ircinia campana), branching-candle sponges ( Ircinia spp.), hard coral (primarily Solenastrea bourgani), miscellaneous sponges, solution holes, and octocorals (sea plumes, primarily Pseudopterogorgia spp.; sea whips, Pterogorgia anceps; and sea rods, primarily Plexaurella spp.). These benthic structures were characterised as “large” (>20 cm diam.) or “small” (<20 cm diam.); structures <20 cm do not shelter juvenile lobster (Childress 1995). For octocorals, only those whose holdfast fell within the belt were counted. Though not living benthos, solution holes large enough (>20 cm) to accommodate juvenile lobster were also counted.

The relative abundance of juvenile lobster was estimated with a 1 h timed search (i.e., catch-per-unit-effort, CPUE). Typically two divers searched separate portions of the survey site for 30 min each. The search time included the time to capture the lobster and record the shelter type used by the lobster. If a capture was missed, divers recorded a shelter type and estimated lobster size. At the end of the search, divers returned the captured lobsters to the boat where carapace length (CL) to the nearest 0.1 mm, sex, moult condition, injuries, and shelter type were recorded. For the 1994 sites, lobsters larger than 50 mm CL were not captured although their
abundance was recorded. For many of the analyses, juvenile lobsters were divided into either three size classes (<25 mm, 25–35 mm, and 35–50 mm CL) or four size classes (15–25 mm CL, 25–35 mm CL, 35–45 mm CL, 45–65 mm CL) similar to other studies (e.g., Herrnkind & Butler 1986; Field & Butler 1994; Childress & Herrnkind 1996).

Data analysis

Descriptive statistics for surface assessments were based on the 355 survey sites in 1994 (surface assessments had not been formalised in 1993), and those for benthic surveys were based on 113 of the 115 benthic survey sites (data from two sites were omitted because of errors in the collection of data). Multivariate analyses for biogeographic characterisations of subregions were based on 108 benthic survey sites covering four subregions with approximately equal sample size (28 sites in Oceanside; 22 sites in Outer Bay; 21 sites in Inner Bay; and 37 sites in Channel). Gulf (1 site) and Basin (4 sites) subregions were retained for descriptive statistics, but were inadequately sampled for reliable multivariate analysis (e.g., the inclusion of the Gulf site would introduce a singularity).

The subregions of the Florida Keys were defined ad hoc based on topography and hydrology categorisations described by Browder (1976) and Schomer & Drew (1982), and confirmed by our own field observations. The Inner Bay comprises a series of shallow basins surrounded by banks that restrict waterflow. Fresh water enters this subregion from the Florida mainland. Freshwater input and rapid evaporation create dramatically changeable salinities in this subregion, particularly in the north (10–60 PSU; Chiappone 1996), however; our surveys were conducted in the southern portion (Fig. 1) where salinities range between 30 and 42 PSU. The Outer Bay subregion covers an open water area north of the middle Keys. This subregion receives little fresh water from the mainland and, lacking the intensive banks of the Inner Bay, water flow between the Gulf of Mexico and Straits of Florida are impeded only by the islands in the middle Keys (Schomer & Drew 1982). Salinities in the Outer Bay thus vary little: typically between 32 and 39 PSU (D’Sa et al. 2001). The Channel subregion (Fig. 1) comprises north-west to southeast channels bordered by islands. These channels range from 14 to 26 km in length by approximately 1 km in width and correspond to a habitat zonation referred to as “high velocity channel” by Browder et al. (1973) or “tidal channels” by Department of Commerce (1996). Rainfall is the only freshwater input to this subregion and salinities are usually between 34 to 36 PSU (D’Sa et al. 2001). The Gulfside and Oceanside subregions are nearshore areas adjacent to the open ocean waters of the Gulf of Mexico and Straits of Florida, respectively. Salinities in these two subregions reflect their open connection to offshore waters and, like the Outer Bay subregion, vary little from 35 PSU. Our use of these subregions also follows the conventions used by Herrnkind et al. (1997) with only minor exception.

Although the subregions described above are generally accepted, their habitat or biogeography have not been quantitatively defined. Therefore, we used a multivariate statistical approach to: (1) characterise subregions within the Florida Keys with respect to general features of hard-bottom habitat that are likely to influence recruitment of juvenile spiny lobster; and (2) examine the association between juvenile lobster abundance and various features of hard-bottom habitat within each biogeographic region. Spatial presentation of data was plotted onto south Florida Arc/Info coverage information provided by the Florida Fish and Wildlife Conservation Commission (FKNMS 2000). For use in these analyses, the raw data from the field surveys were summarised in the following ways. The percentage cover of red macroalgae and seagrass was determined by summing patch sizes for each taxon and dividing by total transect length (200 m in 1993 and 150 m in 1994) for each site. The density of hard-bottom structures was determined by summing the counts of large structures by type and converting them to number per hectare. The three taxa of octocorals were grouped into a single variable owing to inconsistencies in field discrimination of species. The relative abundance of juvenile lobster abundance was defined as the count of lobsters <50 mm CL found within 60 min.

We examined juvenile lobster preference for shelter in the different subregions using a log-likelihood chi-squared test described by Neu et al. (1974). This procedure is robust to scaling changes and Type II errors, and permits the determination of confidence limits for the indices (Allredge & Ratti 1992 a,b; McClean et al. 1998).

Discriminant analysis was used to examine the most parsimonious combination of shelter and lobster size class information that best described the various regions. For the analysis, we used a data matrix that included the density of all types of
structure and the abundance of three size classes of juvenile lobsters using subregion as the grouping variable.

Non-linear canonical correlation analysis was used to examine the relationships of the shelters and lobster size classes (Meulan & Heiser, 2004). The advantages of the non-linear canonical correlation is that it is robust to many zeros, ordinal and nominal variables can both be used in a single analysis, and highly correlated variables can be placed into a single variable set that prevents those correlations from dominating the analysis and masking other relationships (Gifi 1990). Categorical variables were first formulated from lobster abundance by size, the abundance of each type of structure, and percentage cover. Then, an equal membership technique was used to define categories wherein observations were ranked then divided into bins of equal size with the first and last member in a bin defining the bounds of a category. This technique is useful when data contain widely divergent distributions and ranges. The number of individuals in each of three lobster size classes was then placed into a single variable set to examine the differences between the size classes and to reduce the high correlations between these variables. All shelter types were placed into different single variable sets.

RESULTS

Distribution of hard-bottom habitat and juvenile lobsters

Of the 355 surface assessments performed during the spring of 1994 throughout the Florida Keys region, 25% were characterised as hard-bottom, and thus surveyed in detail by divers. This result compared well with a post hoc spatial analysis of the Arc/Info benthic coverages in which 31% of the region was designated as hard-bottom. Within subregions, the comparison between the percentage of hard-bottom from the benthic cover data and percentage from our surface assessments was weaker. For example, in the Channels subregion, the GIS benthic cover data estimated that 29% of this subregion comprised hard-bottom, whereas 44% of our field surveys were over hard-bottom.

The first two discriminant functions described 97% of the variation in the data matrix and comprised the abundance of corals, octocorals, solution holes, three sponge taxa, and three size classes of juvenile lobsters. The first discriminant function was dominated by hard shelter structures (solution holes and coral) plus octocoral densities. The second function comprised vase sponges, miscellaneous sponges, and the three size classes of juvenile lobsters. The configuration of the confidence ellipses (Fig. 2) mirror in many ways the geographical configuration of the ad hoc topographic and hydrological subregions. For example, the subregion Outer Bay occupied an undifferentiated position in the discriminant analysis, suggesting that it constitutes a subset of the Ocean subregion. The Outer Bay shares two broad waterway connections with the Ocean subregion through the Long Key bridge (>3 km) and Seven Mile bridge (10 km). The Channel and Inner Bay subregions are the most differentiated pair of subregions and they are also the most geographically separated. Although discriminant analysis suggests that the Inner Bay and the Channel subregions have the greatest differentiation, they shared a similar distribution of discriminant function 2 scores. All three lobster sizes classes correlated best with discriminant function 2 and both subregions contained the highest juvenile lobster densities.

The abundance of structures of potential importance to lobsters as shelter varied by several orders of magnitude among the four subregions (Fig. 3, Table 1). Octocorals were the most common structure throughout all of the subregions except the Channels. Sponges were most common in the Channels followed by the Outer Bay. Solution holes and coral heads, although not common, reached their highest densities in the Inner Bay. Juvenile lobsters of all sizes were most abundant in the Channels and Inner Bay and least abundant along the Oceanside.

Unlike shelter density, discriminant analysis could not differentiate the subregions based solely on the percentage of different potential settlement substrates (i.e., seagrass, red macroalgae, and green algae) and the abundance of lobsters by size class. Likewise, the non-linear canonical correlation analysis produced non-specific results. For example, when we added one settlement substrate at a time into the shelter and lobster size class model, each loaded positively on the first dimension with the sponges and smallest size class of lobsters. So none of the settlement substrates added any additional discriminatory power exceeding that already attributable to the abundance of hard-bottom structure or lobster size class.

Hard-bottom structure—lobster relationships

Disregarding any differences in shelter use among subregions, a more general analysis of juvenile lobster shelter preference using the entire data set revealed
Table 1  Mean (±SE) density of shelters and mean relative abundance (catch-per-unit-effort) of juvenile lobsters (by size class) for each subregion of the Florida Keys, Florida, United States. Percentages reflect proportions of each shelter or lobster size class within a given subregion.

<table>
<thead>
<tr>
<th>Shelter</th>
<th>Ocean (28)</th>
<th>Outer Bay (22)</th>
<th>Inner Bay (21)</th>
<th>Channel (37)</th>
<th>Basins (4)</th>
<th>Gulf (1)</th>
<th>Overall</th>
</tr>
</thead>
<tbody>
<tr>
<td>Sea whips</td>
<td>2794 ±873</td>
<td>2156 ±443</td>
<td>4857 ±1219</td>
<td>418 ±132</td>
<td>450 ±450</td>
<td>1200 na</td>
<td>2247 ±363</td>
</tr>
<tr>
<td>Loggerhead sponges</td>
<td>333 ±81</td>
<td>283 ±90</td>
<td>278 ±119</td>
<td>598 ±98</td>
<td>433 ±96</td>
<td>133 na</td>
<td>408 ±48</td>
</tr>
<tr>
<td>Misc. sponges</td>
<td>82 ±36</td>
<td>306 ±95</td>
<td>663 ±135</td>
<td>405 ±78</td>
<td>267 ±112</td>
<td>0 na</td>
<td>348 ±46</td>
</tr>
<tr>
<td>Vase sponges</td>
<td>50 ±19</td>
<td>84 ±36</td>
<td>7 ±4</td>
<td>488 ±94</td>
<td>350 ±110</td>
<td>0 na</td>
<td>201 ±37</td>
</tr>
<tr>
<td>Finger sponges</td>
<td>54 ±24</td>
<td>293 ±157</td>
<td>120 ±51</td>
<td>197 ±45</td>
<td>333 ±170</td>
<td>0 na</td>
<td>171 ±37</td>
</tr>
<tr>
<td>Coral</td>
<td>68 ±27</td>
<td>27 ±11</td>
<td>344 ±78</td>
<td>12 ±7</td>
<td>6 ±6</td>
<td>200 na</td>
<td>101 ±20</td>
</tr>
<tr>
<td>Solution holes</td>
<td>25 ±12</td>
<td>25 ±7</td>
<td>174 ±26</td>
<td>9 ±6</td>
<td>6 ±6</td>
<td>0 na</td>
<td>47 ±9</td>
</tr>
<tr>
<td>Total</td>
<td>3406 ±927</td>
<td>3174 ±519</td>
<td>6441 ±1297</td>
<td>2128 ±253</td>
<td>1846 ±426</td>
<td>1533</td>
<td>3523 ±388</td>
</tr>
</tbody>
</table>

Lobster CPUE

<table>
<thead>
<tr>
<th>Size Class</th>
<th>Ocean (28)</th>
<th>Outer Bay (22)</th>
<th>Inner Bay (21)</th>
<th>Channel (37)</th>
<th>Basins (4)</th>
<th>Gulf (1)</th>
<th>Overall</th>
</tr>
</thead>
<tbody>
<tr>
<td>&lt;25 mm CL</td>
<td>0.5 ±0.2</td>
<td>0.9 ±0.3</td>
<td>1.7 ±0.4</td>
<td>3.2 ±0.7</td>
<td>0.5 ±0.5</td>
<td>0 na</td>
<td>1.8 ±0.3</td>
</tr>
<tr>
<td>25–35 mm CL</td>
<td>0.5 ±0.2</td>
<td>1.4 ±0.6</td>
<td>3.8 ±0.8</td>
<td>3.5 ±0.6</td>
<td>1.0 ±0.7</td>
<td>0 na</td>
<td>2.4 ±0.3</td>
</tr>
<tr>
<td>35–50 mm CL</td>
<td>1.0 ±0.3</td>
<td>1.2 ±0.3</td>
<td>3.1 ±0.7</td>
<td>2.6 ±0.5</td>
<td>0.5 ±0.5</td>
<td>0 na</td>
<td>2.0 ±0.2</td>
</tr>
<tr>
<td>Total</td>
<td>2.0 ±0.5</td>
<td>3.5 ±0.8</td>
<td>8.6 ±1.5</td>
<td>9.4 ±1.3</td>
<td>2.0 ±1.7</td>
<td>0 na</td>
<td>6.2 ±0.6</td>
</tr>
</tbody>
</table>
shelter availability was taken into account (Fig. 5). The smallest juvenile lobsters (15–35 mm CL) preferentially used a variety of commercial and non-commercial sponges (including the ubiquitous loggerhead sponge) and avoided hard, rocky substrates such as solutions holes and hard corals. Vase sponges were preferred by juvenile lobsters of intermediate size (35–45 mm CL) and large juvenile lobsters preferred vase sponges, solution holes, and hard coral heads.

Comparison of juvenile lobster shelter preferences among subregions were generally consistent with...
the more general analysis described above, the exceptions being attributable to differences in shelter availability among subregions (Fig. 6). Branching-candle sponges (*Ircinia* sp.) and octocorals were significantly underused across all subregions. In contrast, solution holes were used significantly above availability in all subregions except the Channels, as were loggerhead sponges, which were used significantly above availability in all subregions except the Outer Bay. Large coral heads were only available in the Inner Bay and outer Bay, where large juvenile lobsters invariably used them as shelter in disproportion to their abundance. Miscellaneous sponges comprised a variety of sponge taxa of varying shapes (from encrusting to barrel shaped) and all were underused in all subregions except the Channels. In the Channels subregion, we often found small (<25 mm CL), solitary lobsters in the crevices of large sponges (>20 cm diam.) encrusted by the green sponge *Haliclona viridens*.

Non-linear canonical correlation analysis revealed multivariate relationships between lobster abundance by size class and benthic habitat structure that were consistent with the single-variable shelter preference analyses. The component loadings chart (Fig. 7) revealed two basic groups of shelters used by juvenile lobsters. All sponges loaded positively along the first dimension, whereas the other shelters (e.g., octocorals, corals, and solution holes) loaded negatively. The placement of the size classes of lobsters among the shelter types suggests a transitional relationship. The component loadings for the smallest juvenile lobsters (<25 mm CL) placed them among sponge shelters, whereas component loadings for medium sized juveniles (25–35 mm CL) were intermediate between sponges and hard shelters. The component loading for the largest juveniles (35–50 mm CL) was closer to the hard shelters (e.g., corals, solution holes) (Table 2).

**DISCUSSION**

The results of our rapid assessment survey of the shallow hard-bottom communities of the Florida Keys yields the first quantitative biogeographic description of this ubiquitous, but largely overlooked habitat. Our results also confirm the value of hard-bottom habitat as nurseries for juvenile spiny lobster. We also detail the relationship between the structural features of the habitat and ontogenetic changes in the use of habitat structures as lobsters grow larger.

**Hard-bottom characterisation**

In general, the biogeographic regions that we subjectively defined at the start of the study based on general topographic features were quantitatively distinguishable by the abundance of juvenile lobster and two classes of hard-bottom structures: (1) octocorals and hard structures (hard corals, octocorals, and rocky solution holes); and (2) large sponges (Fig. 2). We found that large sponges were abundant in the Channel subregion, but octocorals and hard structures were not. The Inner Bay subregion was typified by greater numbers of octocorals and hard structures and fewer sponges. The Inner Bay subregion was typified by greater numbers of octocorals and hard structures and fewer sponges. The density of nearly all structures on shallow hard-bottom in the Ocean and Outer Bay subregions were generally low. There were individual sites that deviated from these general patterns and our results do not include hard-bottom areas further offshore that are deeper and more closely associated with the reef tract (see Chiappone & Sullivan 1994b).

One particularly striking biogeographic pattern was the near absence of juvenile lobsters in the Gulfside and Basin subregions of the Florida Keys, which we suspect reflects an inadequate influx of postlarvae to these subregions. In the Basin subregion, a web-like network of emergent mudbanks and small mangrove islands restricts
water flow and offers no direct connection to the sea, the source of lobster postlarvae for settlement. Field & Butler (1994) have shown that a similar situation occurs in the interior of nearby Florida Bay, where mudbanks impede postlarval transport. The Gulfside subregion contains little habitat that is also suitable for lobster settlement and, although open to the Gulf of Mexico to the north, this region is also blocked by islands and mudbanks from the primary source of postlarvae, which arrive from the Straits of Florida to the south of the Florida Keys. No postlarvae have been caught on artificial collectors placed in this subregion to monitor the arrival of postlarvae from the Gulf of Mexico (David Eggleston, North Carolina State University pers. comm.), whereas collectors placed on the oceanside of the Florida Keys are reliable indicators of postlarval supply (Acosta et al. 1997).

**Hard-bottom as a nursery habitat for spiny lobster**

The pattern of shelter preference by juvenile lobsters that we observed in field surveys conducted over a large spatial scale (i.e., the entire Florida Keys region; c. 10,000 km²) and for a wide range of juvenile lobsters sizes (e.g., approximately 15–50 mm CL), is consistent with those predicted from previous experimental investigations. As juvenile
lobsters grew larger, their preference for shelters changed from sponges to hard substrates such as rocks and corals. However, the specific types of shelters used by lobsters in any particular region varied depending primarily on the availability of shelters that make suitable dens. When natural shelters are limited, suboptimal shelters are readily occupied. For example, Childress & Hunt (2002) examined changes in shelter availability and use by juvenile lobster on sites affected by sponge die-offs (Butler et al. 1995) compared with unaffected areas. They concluded that juvenile lobsters were insensitive to changes in the abundance of hard-bottom structure, because lobsters adjusted their use of shelter in proportion to its availability. Although their analysis did not include lobster size, which our study and those of others show is important for shelter selection (Eggelston & Lipcius 1992; Butler & Herrnkind 1997; Lipcius et al. 1998; Lozano-Alvarez & Briones-Fourzan 2001), it nonetheless demonstrates the plasticity of lobster shelter use. Lobsters respond to the sheltering qualities of objects, not the object per se, which is why lobsters will occupy artificial shelters employed for experimental purposes (Butler & Herrnkind 1992, 1997; Lozano-Alvarez et al. 1994), as well as other human-made objects abandoned on the seafloor. Tyres, abandoned lobster traps, oil drums, and other human-made objects housed about 6% of the juvenile lobsters that we observed, indicating the opportunistic nature of shelter use by *P. argus*.

Although lobsters choose many types of objects as dens, certain characteristics of den quality appear to influence their choice. For example, newly settled early benthic juvenile *P. argus* prefer to occupy architecturally complex stands of macroalgae offering a complex network of interstices within which the early benthic juveniles manoeuvre. They choose these shelters over architecturally more simple artificial substrates of similar volume and food quality to natural macroalgae (Herrnkind & Butler 1986; Butler et al. 1997). Experimental studies have shown that the selection of dens by larger juvenile spiny lobsters is driven primarily by size-specific preferences for dens with holes that closely parallel their own dimensions (Eggelston & Lipcius 1992; Lipcius et al. 1998; Lozano-Alvarez & Briones-Fourzan 2001). However, their selection of dens is modified by the chemically attractive cue of conspecifics (Zimmer-Faust et al. 1985; Childress & Herrnkind 1997, 2001; Ratchford & Eggelston 1998; Butler et al. 1999; Nevitt et al. 2000; Lozano-Alvarez & Briones-Fourzan 2001) and their repulsion for diseased conspecifics (Behringer et al. 2006) and certain predators, especially octopus (Berger & Butler 2001). Predation itself further modifies the resultant patterns of den occupancy, if not den choice (Eggleston et al. 1990, 1997; Schratwieser 1999).

Although the relationship between the abundance of large juvenile lobsters (>35 mm CL) and the availability of benthic structures is clear, discerning similar patterns for more recently settled juveniles has been elusive. We could not show a strong link between settlement habitat abundance (i.e., red macroalgae) and the abundance of juvenile lobsters (large or small) using data from our Florida Keys-wide surveys, corroborating previous studies conducted at smaller spatial scales. Field & Butler (1994) studied shelter use of post-algal juveniles (15–60 mm CL) within the Inner Bay subregion and examined the relationship between lobster abundance and the density of various structures and settlement substrates with multiple regression. They found that the abundance of larger (>35 mm CL) juvenile lobsters could be predicted by salinity and the density of octocorals, corals, and solution holes. However, they also could not find a predictable relationship between macroalgae and lobsters less than 35 mm CL (Field & Butler 1994). Herrnkind & Butler (1994) also used multiple regression to examine the relationship between juvenile lobster abundance (by size class) within the Inner Bay subregion and a suite of measures, including postlarval supply, macroalgal coverage, and the density of different hard-bottom structures. They found that the best predictor for all lobster size classes was the abundance of sponges (Herrnkind & Butler 1994). The lack of a strong link between macroalgal abundance and juvenile lobsters may be because of the ephemeral nature of the distribution of the macroalgae (Mathieson & Dawes 1975, Butler et al. 1997), or our inability to effectively survey lobsters during their “transitional phase”, when they are cryptic, solitary, and transient between macroalgae and benthic structures (Childress & Herrnkind 2001).

**Hard-bottom monitoring and protection**

One aspect of this study, the regional characterisation of shallow hard-bottom habitat, parallels similar efforts to describe the biogeography of two other major marine habitats in the Florida Keys, namely seagrass (Zieman et al. 1989, 1999) and coral reefs (Florida Fish and Wildlife Conservation Commission & National Oceanic and Atmospheric Administration 2000; Lidz et al. 2006). Our focus in this study was to determine which aspects of hard-bottom
habitat were most relevant as a nursery role for lobsters. Thus, the level of taxonomic detail we used was admittedly coarser and used fewer types of measures of community structure than those used in more detailed studies of fewer sites (Chiappone & Sullivan 1994a,b). However, this study constituted a first step in hard-bottom characterisation at a time when GIS depictions of habitat coverage were unavailable, and it allowed us to describe general biogeographic patterns in hard-bottom structure by surveying many sites (c. 100) in a short amount of time (c. 1 month)—the very goal of rapid assessment procedures, especially those potentially used in developing nations where GIS-based maps of subtidal habitats may be unavailable.

This project also served to establish a long-term monitoring programme for shallow hard-bottom habitat in the Florida Keys. Like most coastal habitats (Finkle & Charlier 2003; Sale et al. 2008), hard-bottom is threatened by a variety of environmental perturbations, most of which are anthropogenic in origin. Impacts on the abundance of lobster have been documented on hard-bottom areas in Florida affected by siltation (Herrnkind et al. 1988). Declining water quality and harmful algal blooms are the suspected culprits that have triggered massive die-offs of sponges in the Florida Keys and the wholesale reconfiguration of impacted hard-bottom areas (Butler et al. 1995; Philips et al. 1999). Commercial fishing of sponges removes approximately 7 million sponges from shallow hard-bottom habitat in the Florida Keys each year (Witzel 1999) with unknown consequences, but the indirect effects of other fishing activities may be more severe. For example, commercial trawling for bait shrimp (Eldred et al. 1972; Berkeley et al. 1985) and the use of traps for lobster and crabs (Matthews et al. 2005) damages hard-bottom areas. Lost fishing gear accounts for over 80% of the debris found along the Florida Keys reef tract and associated offshore hard-bottom areas, and it causes over 80% of the physical damage to sessile invertebrates (Chiappone et al. 2005). Damage to juvenile lobsters dwelling in hard-bottom and caught incidentally by commercial and recreational fishers substantially diminishes the recruitment of lobster to the Florida Keys fishery (Lyons & Kennedy 1981; Hunt & Lyons 1986; Forcucci et al. 1994). Compounding these anthropogenic disturbances to hard-bottom habitat are the adverse effects of episodic natural disturbance caused by hurricanes. In short, the importance of hard-bottom habitat in south Florida for economically important species such as spiny lobster, juxtaposed against the suite of human activities that impinge on its sustainability, provides a compelling case for adding shallow hard-bottom to the list of marine habitats deserving more vigilant, regular monitoring and protection.

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